

# Investigating Surficial Geologic Controls on Soil Properties, Inorganic Nutrient Uptake, and Northern Hardwood Growth in Western Massachusetts, USA

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## Abstract

The influence of glacial geologic materials on soil properties, tree nutrient acquisition, and tree growth rates in New England is not well-constrained. Here, our research investigates the effect of two dominant surficial deposits, glacial till and glaciofluvial deposits, on soils and northern hardwood trees in western Massachusetts. We investigated sixteen paired glaciofluvial and glacial till sites located on the perimeters of glacial lake Hitchcock sediments, which drained 12,400 years ago. At each site, a 12.2-m-radius circular plot was selected, a soil pit was excavated, and all trees within the plot were sampled for foliage and cored with an increment borer. Our analyses found glaciofluvial soils to have significantly higher pH, clay fraction, and water field capacity than glacial till soils. Glaciofluvial soils also had less rock fragments and lower sand content than glacial till soils. We observed significantly higher pseudo-total K, Ca, Mg, and Mn concentrations in glacial till soils, but found similar foliar concentrations for five of the six tree species. Tree cores showed Black Birch, Red Maple, and Red Oak grew 1.3 to 2.1 times faster on glaciofluvial soils. Our study found that glaciofluvial soils, which exhibit greater water retention, less rocks, more fine particles, and higher soil pH than glacial till soils, promote faster growth of Black Birch, Red Maple, and Red Oak. However, the growth of American Beech, White Oak, and Eastern Hemlock was not impacted by surficial deposits, implying adaptation to nutrient limitations, coarser rocky soils, and potential water stress. Thus, the growth of some common tree species is affected by geologic materials, but others are not affected.

**Keywords** Glacial till · Glaciofluvial · Soil parent material · Forest nutrients

## 1 Introduction

In the northeastern USA, forests are critical resources for harvesting timber, building materials, and biomass as biofuel for domestic heating. While the demand for timber and biofuel is increasing, forest lands dedicated to timber harvesting management plans are simultaneously decreasing (Robertson et al. 2011; Joshi and Mehmood, 2011). Thus, it is increasingly important for public and private forest and land managers to consider the longevity of nutrients and soil properties for sustainability of production rates of managed forested lands (Deal et al. 2012; Paré and Thiffault, 2016). Soil fertility is central to the sustainability of forestry, but few studies have explored the

link between soil properties to their geologic parent materials in the northeastern USA (e.g., Li et al. 2017). Here, we investigate the link between geologic materials and physical and chemical properties of soils, and their subsequent influence on nutrient uptake and tree growth.

Glacial geology dominates the formation of soils in the northeastern USA, including the state of Massachusetts. Most of Massachusetts' soils derive from either glaciofluvial or glacial till deposits. During the last glacial maximum from 19,000–22,000 years ago, the Laurentide Ice Sheet covered Massachusetts (Dyke and Prest, 1987; Dyke et al. 2002). As the ice sheet moved southward, it deposited a heterogeneous mix of rock fragments, ranging from sand to boulders > 4 m diameter, forming lodgement glacial till (Dyke and Prest, 1987). Glacial till is the most common geologic material on upslope, elevated topographic positions in the northeastern USA. The Laurentide Ice Sheet retreated and glacial till-derived soils began forming approximately 19,000 to 15,000 years ago in Connecticut and Massachusetts (Balco

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and Schaefer 2006) or as late as 11,000 years ago in Vermont and New Hampshire (Ridge and Larsen, 1990), as determined by studies of varved lake sediments and cosmogenic radionuclide dating. Fluvial systems weathered, eroded, and re-deposited the glacial till-derived soils as glaciofluvial deltaic and lacustrine deposits. In the Connecticut River Valley of western Massachusetts, much of the deltaic and lacustrine deposits correspond with Glacial Lake Hitchcock sediments deposited between 16,500 and 12,400 years ago (Dyke and Prest, 1987; Ridge and Larsen, 1990; Uchupi et al. 2001).

Surficial geology can control many edaphic properties of their overlying soils. Glaciofluvial deposits are typically well-sorted and dominated by fine sands and coarse silt while lacustrine deposits, particularly the varved materials, may be dominated with fine silts and clays (Hartshorn and Young, 1969; Ashley, 1975). Soils formed from poorly sorted glacial deposits have higher boulder, stone, and gravel content. The sandy loam, rocky glacial till soils of western Massachusetts are excessively well-drained while silt loams formed from finer glaciofluvial deposits are commonly poorly drained (Villholth et al. 1998; Soil Survey Staff, 2008). These differences have important implications for tree growth as excessively drained soils are drier during late-summer droughts, but finer soil textures can decrease oxygen diffusivity to roots and increase surface area for weathering and sorption (Mohanty and Mousli 2000; Taylor and Blum 1995; Miller et al. 1993). Moreover, there may be mineralogical differences in glacial till and glaciofluvial deposits, due to the heterogeneity of material deposited by the Laurentide Ice Sheet (Bailey and Hornbeck 1992) or preferential losses of carbonates and increases in feldspars and quartz during sediment transport (Eberl 2004).

Soils derived from glacial till and glaciofluvial deposits may influence tree growth differently due to differences in their reservoirs of inorganic nutrients and the release rates of these nutrients. Paoli et al. (2007) found that surface soil P, K, and Mg concentrations and percent sand content were significantly related to forest stem density and aboveground biomass in a tropical forest. Paoli et al. (2007) also determined that 31% of the aboveground biomass variance observed was due to P availability and percent sand content. As another example, Royer-Tardif and Bradley (2011) found evidence that soil nutrient availability controlled the relative abundance of Jack Pine and Trembling Aspen in Quebec, Canada. They also found that fertile clay deposits fostered a more heterogeneous forest tree species composition than on nutrient-poor glacial till sites (Royer-Tardif and Bradley, 2011). Calvaruso et al. (2017) found that soils underlying European Beech (*Fagus sylvatica*) stands inherited key physical and chemical soil properties from their geologic materials. Analysis of plant data from the U.S. Forest Service “Tree Chemistry Database” revealed differences in foliar chemistry based on the geologic materials (Pardo et al. 2005); fluvial K ( $6.2 \pm 0.1 \text{ mg g}^{-1}$ ) and

Ca ( $7.5 \pm 0.1 \text{ mg g}^{-1}$ ) concentrations were lower than glacial-till K ( $7.7 \pm 0.1 \text{ mg g}^{-1}$ ) and Ca ( $9.2 \pm 0.1 \text{ mg g}^{-1}$ ) concentrations, respectively. However, few studies have examined the impact of geologic materials on tree growth on the northern hardwoods that dominate the forests of New England (Finzi et al. 1998).

The primary objective of this study was to explore the influence of geologic materials (glaciofluvial and glacial-till) on soil properties and assess if they significantly affect tree nutrient acquisition and growth rates in unmanaged northern hardwood forests of western Massachusetts. Our first hypothesis was that glaciofluvial soils would promote greater tree growth than glacial till soils due to their physical properties (less rocks, higher water field capacity, more clay) and chemical properties (higher pH, higher concentrations of Ca, K, Mg, Mn, Cu, Zn). Our second hypothesis was that glaciofluvial soils would promote greater tree nutrient uptake and growth rates than glacial till soils due to their edaphic properties (higher pH, greater water field capacity, higher nutrient availability). The information may be useful for forest ecosystem researchers and forest resource managers to determine differences in site productivity influenced by surficial geology.

## 2 Materials and Methods

### 2.1 Site Descriptions

We studied 16 paired forested sites across the two dominant surficial deposits in the Connecticut River Valley of western Massachusetts (Table 1, Fig. 1). Each pair consisted of a glaciofluvial and glacial till forest site, within 400 m of each other, on the edge of Glacial Lake Hitchcock lacustrine deposits and Wisconsinian glacial till. Each pair of forest sites had comparable elevation, aspect, and geomorphic position. The glacial material for each site was first identified using the United States Department of Agriculture Natural Resource Conservation Service’s web tool Web Soil Survey (<https://websoilsurvey.sc.egov.usda.gov/App/HomePage.htm>, accessed June 2018) and the USGS 1:24,000 Surficial Geology Map (<https://docs.digital.mass.gov/dataset/massgis-data-usgs-124000-surficial-geology>, accessed June 2018) and further confirmed through soil pit excavation. The location of each forest site is detailed in (Table 1, Fig. 1). Each potential pair was inspected for forest composition, geologic material, and hydrology. Forests with human disturbances, recent forest management activities, boulder fields, exotic tree species, and poor drainage were deemed unsuitable for this study. Forest sites also needed to be well-drained, on slopes  $< 10^\circ$ , and at least 50 m from any human roads or structures. At forest sites deemed suitable, we denoted a circular forest stand with a 12.2-m (40 ft) diameter for study.

**Table 1** Location of forest stands, their tree species composition, and stand age estimate

t1.2	Site #	Material	Soil series	Latitude	Longitude	Elevation a.s.l. (m)	Tree species <sup>†</sup>	Stand age <sup>‡</sup> (years)
t1.3	1	Glaciofluvial	Belgrade	42° 17' 16.98" N	72° 32' 39.55" W	93	AB, WO	44 ± 5
t1.4		Glacial Till	Narragansett-Holyoke	42° 17' 27.20" N	72° 32' 44.66" W	110	RM, WO	47 ± 5
t1.5	2	Glaciofluvial	Belgrade	42° 17' 9.42" N	72° 34' 34.83" W	51	EH, RO	58 ± 8
t1.6		Glacial Till	Holyoke	42° 17' 21.95" N	72° 34' 36.26" W	93	EH, RO	48 ± 11
t1.7	3	Glaciofluvial	Hinckley	42° 17' 2.26" N	72° 36' 1.01" W	49	BB, EH, RM, RO	71 ± 21
t1.8		Glacial Till	Holyoke	42° 17' 4.07" N	72° 36' 8.34" W	47	BB, EH, RM, WO	60 ± 5
t1.9	4	Glaciofluvial	Hinckley	42° 18' 25.85" N	72° 34' 23.12" W	171	BB, EH, RM	67 ± 15
t1.10		Glacial Till	Narragansett-Holyoke	42° 18' 20.81" N	72° 34' 21.68" W	203	BB, EH, RM	55 ± 12
t1.11	5	Glaciofluvial	Hinckley	42° 18' 38.27" N	72° 33' 22.97" W	138	AB, EH	78 ± 13
t1.12		Glacial Till	Narragansett-Holyoke	42° 18' 40.39" N	72° 33' 19.04" W	147	AB, EH, RO	56 ± 11
t1.13	6	Glaciofluvial	Belgrade	42° 17' 41.50" N	72° 39' 31.86" W	64	EH, RM	67 ± 17
t1.14		Glacial Till	Narragansett-Holyoke	42° 17' 53.63" N	72° 39' 37.30" W	95	BB, WO	58 ± 10
t1.15	7	Glaciofluvial	Hinckley	42° 16' 20.90" N	72° 36' 44.86" W	55	BB, EH, RM	78 ± 19
t1.16		Glacial Till	Narragansett-Holyoke	42° 16' 20.21" N	72° 36' 49.68" W	56	AB, RM	68 ± 20
t1.17	8	Glaciofluvial	Hinckley	42° 14' 21.98" N	72° 39' 28.40" W	76	BB, RM, WO	48 ± 12
t1.18		Glacial Till	Narragansett-Holyoke	42° 14' 13.45" N	72° 39' 16.52" W	150	BB, WO	64 ± 21
t1.19	9	Glaciofluvial	Hinckley	42° 14' 4.70" N	72° 39' 27.68" W	117	BB, WO	81 ± 17
t1.20		Glacial Till	Narragansett-Holyoke	42° 14' 1.14" N	72° 39' 25.74" W	151	AB, BB	56 ± 13
t1.21	10	Glaciofluvial	Hinckley	42° 21' 2.74" N	72° 39' 22.14" W	82	BB, EH	69 ± 22
t1.22		Glacial Till	Charlton-Rock outcrop-Hollis	42° 21' 5.36" N	72° 39' 28.69" W	97	BB, EH	68 ± 23
t1.23	11	Glaciofluvial	Belgrade	42° 21' 14.00" N	72° 40' 5.22" W	74	BB, EH, RM, WO	41 ± 9
t1.24		Glacial Till	Charlton-Rock outcrop-Hollis	42° 21' 22.00" N	72° 40' 2.75" W	80	EH, BB, RM, WO	39 ± 9
t1.25	12	Glaciofluvial	Belgrade	42° 22' 52.24" N	72° 38' 39.92" W	58	EH, BB, WO	60 ± 13
t1.26		Glacial Till	Charlton-Rock outcrop-Hollis	42° 22' 52.89" N	72° 38' 36.97" W	67	EH, BB, RM, WO	81 ± 20
t1.27	13	Glaciofluvial	Hinckley	42° 22' 27.77" N	72° 39' 50.47" W	86	AB, EH, RO	88 ± 16
t1.28		Glacial Till	Woodbridge	42° 22' 32.67" N	72° 39' 56.88" W	113	AB, EH, RO, BB	85 ± 16
t1.29	14	Glaciofluvial	Hinckley	42° 27' 55.12" N	72° 30' 11.74" W	149	BB, EH, RO	77 ± 15
t1.30		Glacial Till	Holyoke-Yalesville	42° 27' 50.40" N	72° 30' 18.11" W	148	AB, EH	70 ± 13
t1.31	15	Glaciofluvial	Hinckley	42° 29' 59.31" N	72° 31' 47.28" W	133	AB, BB, RO	68 ± 11
t1.32		Glacial Till	Holyoke-Yalesville	42° 29' 57.16" N	72° 31' 49.80" W	145	BB, EH, RO	73 ± 14
t1.33	16	Glaciofluvial	Scio	42° 32' 57.30" N	72° 34' 28.67" W	81	BB, EH, RO	80 ± 12
t1.34		Glacial Till	Holyoke-Yalesville	42° 33' 5.29" N	72° 34' 31.37" W	116	AB, EH, RO	71 ± 17

<sup>†</sup> Forest species codes: AB = American Beech, BB = Black Birch, EH = Eastern Hemlock, RM = Red Maple, RO = Red Oak, WO = White Oak

<sup>‡</sup> Stand age was estimated to be the average tree age using counted tree core rings

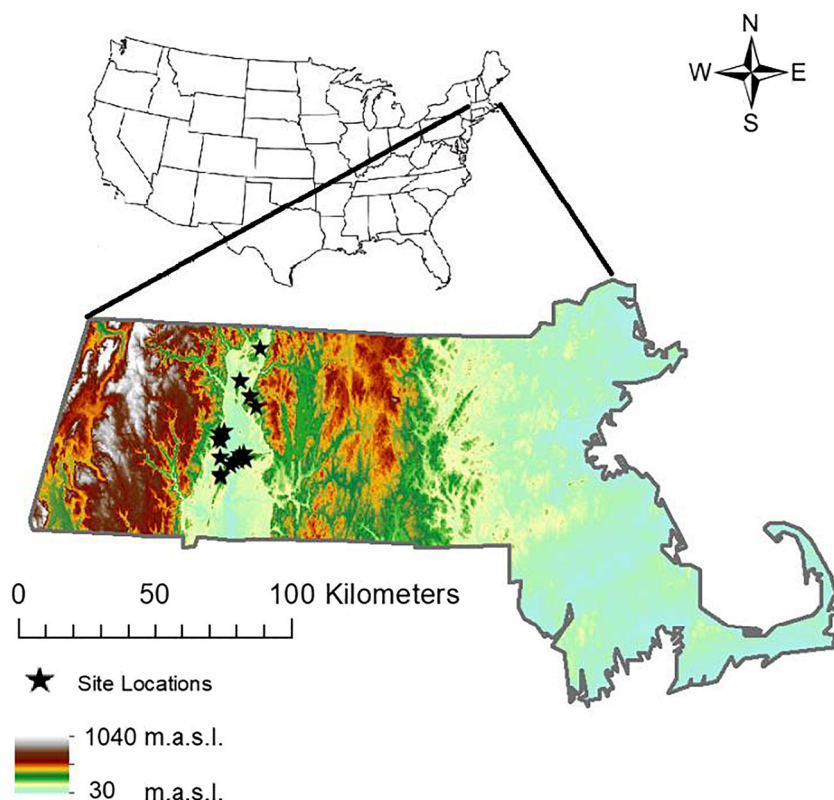
162 Current land management of all sites is single-tree selection  
163 harvesting and reforestation conservation.

## 164 2.2 Tree Species, Tree Coring, and Foliar Analyses

165 Tree species present at glacial till and glaciofluvial sites are  
166 given in Table 1. All trees with > 10 cm diameter were mea-  
167 sured for diameter at breast height (DBH) and identified. Leaf  
168 foliage and bolewood samples were collected from American  
169 Beech (*Fagus americana*), Black Birch (*Betula lenta*), Eastern  
170 Hemlock (*Tsuga canadensis*), Red Maple (*Acer rubrum*),

White Oak (*Quercus alba*), Red Oak (*Quercus rubra*), 171  
White or "paper" Birch (*Betula papyrifera*), and White Ash 172  
(*Fraxinus americana*) in triplicate from each forest stand in 173  
the summer of 2018. Species were identified using the *Trees of* 174  
*Eastern North America* dichotomous tree guide (Nelson et al. 175  
2014). Foliage was collected from branches in the middle to 176  
upper canopy, 4 to 25 m above the ground, using a stainless- 177  
steel pole saw (see Richardson and Friedland, 2016) or an 178  
arborist throw-ball. For the throw-ball technique, a 0.4-kg 179  
arborist throw-ball was lobbed over upper canopy branches 180  
and the branches were forcibly removed at the connection to 181

**Fig. 1** Location of sampling sites indicated by stars across the state of Massachusetts, USA, with meters above sea level (m.a.s.l.) from a digital elevation map. The Connecticut River Valley closely corresponds with glacial lake Hitchcock and former fluvial sediments



182 the main trunk. For shorter trees, a stainless steel pole saw was  
183 extended, and a branch was collected from the main trunk. In  
184 both cases, branches collected were between 3 and 10 cm in  
185 diameter.

186 Trees were cored using an 18.3-cm-long increment corer  
187 with a 4.3-mm width. Trees with a DBH of 10 cm and under  
188 were not cored. In the laboratory, all cores were secured to a  
189 30-cm-long wood board with glue. Tree cores were polished  
190 starting with sandpaper at 40 grit and continued down to 800  
191 grit. The polished tree cores were analyzed by dissecting mi-  
192 croscope to count and measure each annual tree ring. Using  
193 the tree ring data, we estimated the age of the tree, annual  
194 growth rate, and minimum stand age (Table 1).

195 To determine macro- and micronutrient concentrations, di-  
196 gestions were carried out using a modified EPA 3050B  
197 Method (Chen and Ma, 1998; Rechcigl and Payne, 1990), in  
198 which samples are combusted prior to strong acid, pseudo-  
199 total digestion. To begin the process, plant material was dried  
200 to a constant weight at 90 to 105 °C for a period of 24 h.  
201 Foliage was then ground up to reduce heterogeneity, and for  
202 larger leaves, the mid-vein was removed prior to grinding. The  
203 ground-up foliage was then transferred to a ceramic vessel and  
204 combusted at 550 °C for 8 h. The ashes were transferred to  
205 50-mL centrifuge tubes and digested with 5 mL of reverse  
206 aqua regia (9:1 HNO<sub>3</sub>:HCl) and lightly capped to degas over-  
207 night. After 12 h, the digest was diluted to 50 g using deion-  
208 ized water. Samples were further treated by diluting 3 g of the

plant tissue digest to 15 g using 2.5% HNO<sub>3</sub> solution for 209  
analysis by inductively coupled plasma-mass spectrometry 210  
(ICP-MS). While this method can cause issues for measuring 211  
insoluble, high field strength elements (e.g., Ti or Si), it is 212  
effective for measuring base cations and micronutrient trace 213  
elements (see Rechcigl and Payne, 1990). 214

### 2.3 Soil Sampling and Analysis 215

216 Soils at each site were sampled between June and August 217  
2018. A 1-m-wide by 1-m-deep soil pit was excavated in the 218  
center of each forest stand. One side of the pit was designated 219  
for pedon description, to avoid any compaction, and described 220  
following U.S. Soil Taxonomy using the National Soil Survey 221  
Center NRCS USDA Field Book for Describing and 222  
Sampling Soils Version 3.0 (Schoeneberger et al. 2012). 223  
Starting from the bottom of the pit, soil cores were collected 224  
from each of the horizons using a 15-cm-diameter steel cylin- 225  
der to capture soil bulk density. For soil horizons with large 226  
rocks (> 5 cm) that prevented the steel cylinder from being 227  
hammered into the soil pit face, the % rock volume was visu- 228  
ally estimated with a 15 × 15 grid, and collected by trowel. 229  
This visual estimation of rock fraction occurred for 20 of the 230  
148 total soil horizons and exclusively for glacial till soils. 231  
One soil sample was obtained from each soil horizon and 232  
collected in polyethylene bags. In total, 32 soil pits were ex- 233  
cavated in this study and 148 soil horizons sampled.

234 All mineral soil samples and organic horizons were air-  
 235 dried. Mineral soil samples were then weighed and sieved to  
 236  $\leq 2$  mm and then re-weighed. A 2:5 soil–water slurry was used  
 237 to determine soil pH. Slurries were shaken for 1 h using a  
 238 wrist-action shaker and vacuum extracted through a  
 239 Whatman 40 filter. The pH of the supernatant extract was  
 240 measured with a pH meter (8015 VWR). For organic-rich  
 241 horizons, samples were filtered using a Whatman 1 filter.  
 242 Loss on ignition was used to estimate % soil organic matter  
 243 (SOM) and measured by combusting a 4-g oven-dried sub-  
 244 sample at 550 °C for 8 h. Every 20 samples included one  
 245 blank and duplicate. To determine the soil particle size distri-  
 246 bution, we weighed  $\sim 30$  g of dried soil into a 250-mL glass  
 247 beaker. Organic matter was removed and we added 100 mL of  
 248 1 M sodium hexametaphosphate (HMP) solution to the soil  
 249 for at least 8 h to disperse soil particles. This HMP–soil slurry  
 250 was washed out into a 1000-mL graduated cylinder with DI  
 251 water. We utilized a modified Bouyoucos hydrometer method  
 252 with hydrometer readings at 60 s and 1.5 h after mixing to the  
 253 closest  $0.5 \text{ g L}^{-1}$  (Gee and Bauder, 1986). To examine differ-  
 254 ences in soil water retention of the soils, we performed a field  
 255 capacity test (Rawls et al. 1991). We performed the test by  
 256 adding 20 g of soil to a funnel with Whatman 1 filter paper,  
 257 then saturating the soil with DI water. Once all the water had  
 258 passed through the sample, we weighed the wet soil at field  
 259 capacity following cessation of gravimetric water draining and  
 260 compared the wet mass to the soil's dry mass to determine the  
 261 percent field capacity. Field capacity is reported as a percent-  
 262 age of the wet weight divided by the soil dry weight.

263 For macro- and microelement analysis, soil samples were  
 264 dried at 105 °C for 24 h and 0.5 g was weighed into 50-mL  
 265 centrifuge tubes for acid digestion. Soils were not ground to  
 266 avoid creating fresh surfaces for dissolution of silicate min-  
 267 erals. Soil samples were extracted using a strong acid, pseudo-  
 268 total digestion with 5 mL of 9:1  $\text{HNO}_3$ :HCl acid heated to  
 269 90 °C for 45 min. This method allows for quantification of  
 270 metals that are sorbed to organic matter and secondary Al and  
 271 Fe oxides not within crystalline silicates, providing an esti-  
 272 mate of metals that are bioavailable or mobile (Chen and  
 273 Ma, 1998). With every 20 samples, a preparation blank, a  
 274 duplicate, and a standard reference material (SRM) was in-  
 275 cluded. Montana Soil 2711a and Peach Leaves 1547a from  
 276 the National Institute of Standards and Technology (NIST)  
 277 were used as SRMs for soil and plant samples, respectively.  
 278 The digests were then diluted to 50 mL using deionized water.

## 279 2.4 ICP-MS Analyses

280 Soil extracts and plant digests were diluted with deionized  
 281 water and analyzed for macro- and micronutrients (Ca, K,  
 282 Mg, Fe, Mn, Cu, Zn) with an Agilent 7700x Inductively  
 283 Coupled Plasma Mass Spectrometer (Agilent Technology,  
 284 Santa Clara, CA, USA). Recoveries for pseudo-total digests

of Ca, K, Mg, Mn, Cu, and Zn were 82–111% of their certified 285  
 values. The metal concentration coefficient of variation be- 286  
 tween intra-sample duplicates was  $< 7\%$ , and metal concen- 287  
 trations in the preparation blank samples were  $< 0.1\%$  of their 288  
 analyte concentrations. 289

## 290 2.5 Statistical Analyses

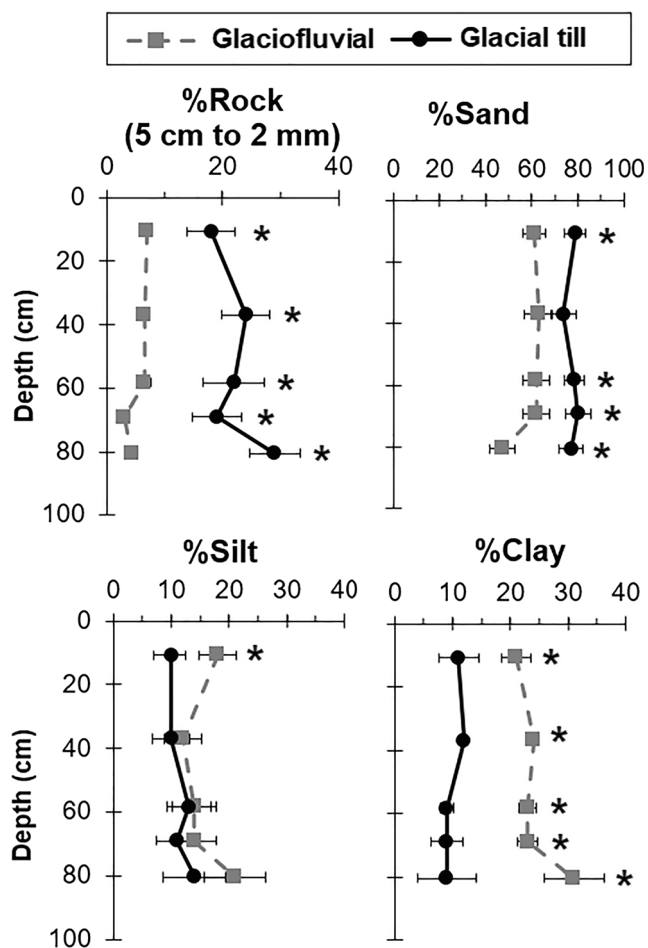
Descriptive statistics were calculated in Matlab (Mathworks, 291  
 Natick, MA, USA). Average values are presented in text and 292  
 in figures  $\pm 1$  standard error. Data were tested for normal dis- 293  
 tribution using the Kolmogorov-Smirnov test (Lilliefors, 294  
 1967) and logarithmically transformed when necessary to es- 295  
 tablish normality. Foliar and mineral soil macro- and micro- 296  
 nutrient concentrations and soil properties (pH, %LOI, field 297  
 capacity, bulk density) were compared between glaciofluvial 298  
 and glacial till sites for both soil horizon and tree species using 299  
 paired sample *t* tests. Differences among tree species were 300  
 determined using two-way analysis of variation (ANOVA) 301  
 tests with post hoc *t* tests. For tree cores, time-series analysis 302  
 was performed using ANTEVS 1.4.1 software to determine 303  
 the detrended averages (Rayburn and Vollmer, 2013). 304

## 305 3 Results

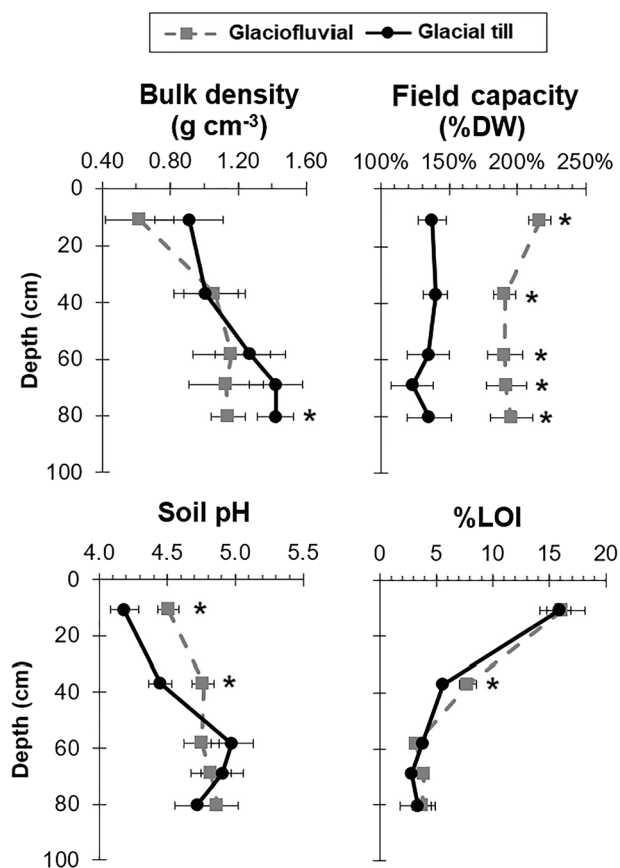
### 306 3.1 Soil Physicochemical Properties Across Glacial Till- 307 and Glaciofluvial-Derived Soils

Our measurements across the 16 paired sites show that there 308  
 are significant differences in soil physical and chemical prop- 309  
 erties between soils derived from glaciofluvial materials and 310  
 soils derived from glacial till materials. According to U.S. Soil 311  
 Taxonomic information from Web Soil Survey and field ob- 312  
 servations, glacial till soils were all Dystrudepts, primarily of 313  
 the Narragansett, Holyoke, Charlton, Hollis, and Yalesville 314  
 soil series, while glaciofluvial soils were predominantly of 315  
 the Belgrade, Hinckley, and Scio soil series. Rock fraction, 316  
 %sand, %silt, and %clay were significantly different between 317  
 glaciofluvial-derived and glacial till-derived soils for each soil 318  
 horizon across the forest sites (Fig. 2). Glacial till soils had a 319  
 significantly higher rock fraction (2 mm to 5 cm in diameter) 320  
 and %sand than glaciofluvial soils for all horizons by  $> 10\%$  321  
 w/w (Fig. 2). Furthermore, the glacial till soils had a signifi- 322  
 cantly lower clay fraction than glaciofluvial soils for all hori- 323  
 zons (Fig. 2). The %silt fraction in the A horizons (0 to 30 cm 324  
 depth) was significantly lower for the glacial till soils com- 325  
 pared with glaciofluvial soils, but not for the other horizons. 326

Surface soil horizon bulk density did not differ between 327  
 surficial geologic materials; however, the C horizons of glacial 328  
 till soils were significantly more dense than the C horizons of 329  
 glaciofluvial soils ( $P < 0.05$ , Fig. 3). Field capacity was sig- 330  
 nificantly greater for glaciofluvial soils compared to glacial till 331



**Fig. 2** Average rock and soil particle size distributions for each horizon at glaciofluvial and glacial till. Error bars are  $\pm 1$  standard error. (\*) indicates a significant difference ( $P < 0.05$ ) using paired  $t$  test



**Fig. 3** Average bulk density, field capacity, soil pH (1:2.5 soil to water ratio), and loss on ignition (%LOI) for each horizon at glaciofluvial and glacial till. Error bars are  $\pm 1$  standard error. (\*) indicates a significant difference ( $P < 0.05$ ) using paired  $t$  test

332 soils for all soil horizons ( $P < 0.05$ , Fig. 3). A horizon and  
 333 upper B horizon soil pH was significantly higher for  
 334 glaciofluvial soils compared to glacial till sites (Fig. 3).  
 335 However, %LOI in glaciofluvial soils was only greater than  
 336 %LOI in glacial till soils in the upper B horizons (~40 cm  
 337 depth) and lower B horizons (~70 cm depth) (Fig. 3).

338 Nutrient concentrations were determined using a pseudo-  
 339 total digestion, allowing for measurement of nutrients that are  
 340 readily plant available or non-crystalline silicate forms that  
 341 may become plant available (e.g., organic complexed, sorbed  
 342 to secondary oxides) (see Chen and Ma, 1998). Using the  
 343 pseudo-total extractions, we found that glacial till soils had  
 344 significantly higher concentrations of Ca, K, Mg, Mn, and  
 345 Zn in their A and upper B horizons (Fig. 4).

### 346 3.2 Tree Nutrient Uptake and Growth Rates

347 Mid-season foliage was collected and analyzed for inorganic  
 348 nutrients to determine if nutrient acquisition differed between  
 349 glaciofluvial and glacial till deposit sites (Fig. 5). A two-way  
 350 ANOVA and post hoc  $t$  tests ( $P > 0.10$ ) revealed no significant

relationship between surficial geology and nutrient acquisition. 351  
 However, our data did show that Red Maple exhibited 352  
 significantly higher foliar Ca, K, and Cu concentrations 353  
 ( $P < 0.05$ , Fig. 5) while growing on glacial till. 354

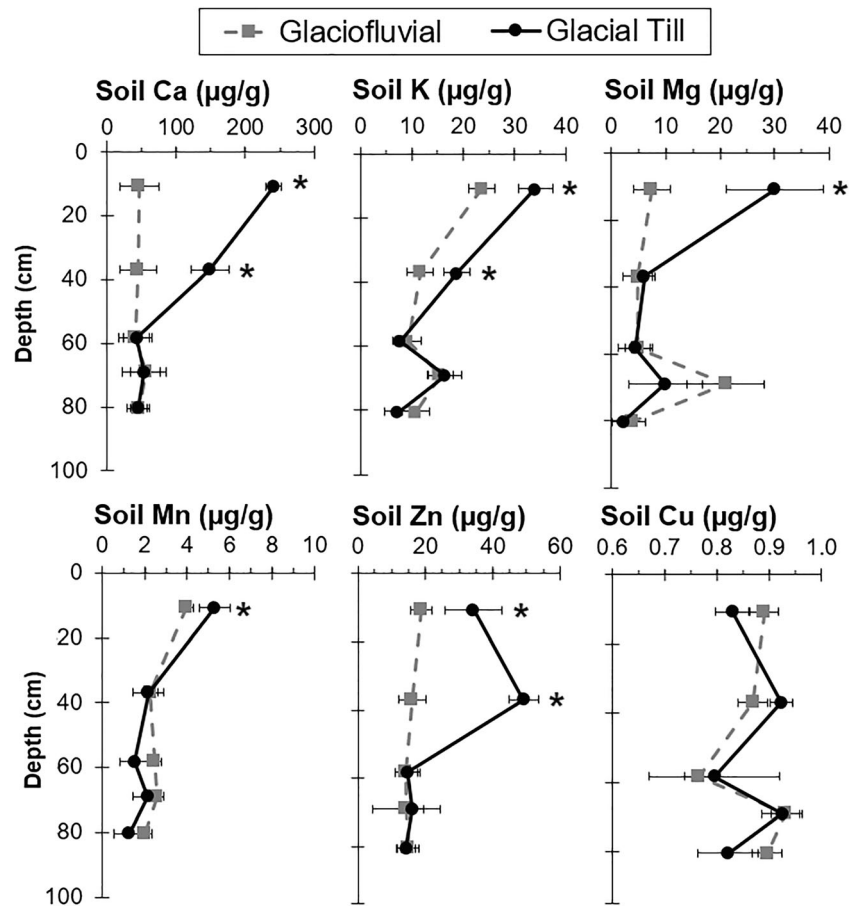
We examined tree age and annual tree growth using tree 355  
 cores. To better quantify non-site-specific effects of growth rate, 356  
 data were grouped among geologic materials and averaged by 357  
 calendar year for each tree species and compared across surficial 358  
 geologic material. From our data shown in Fig. 6, Black 359  
 Birch, Red Maple, and Red Oak had faster annual growth rates 360  
 on the glaciofluvial soils than glacial till soils. However, 361  
 American Beech, Eastern Hemlock, and White Oak grew at 362  
 similar rates for both surficial geologic materials. 363

## 364 4 Discussion

### 365 4.1 Soil Physicochemical Properties Between Glacial 366 Till- and Glaciofluvial-Derived Soils

Our analyses support our first hypothesis that soils derived 367  
 from glacial till and glaciofluvial geologic materials have 368

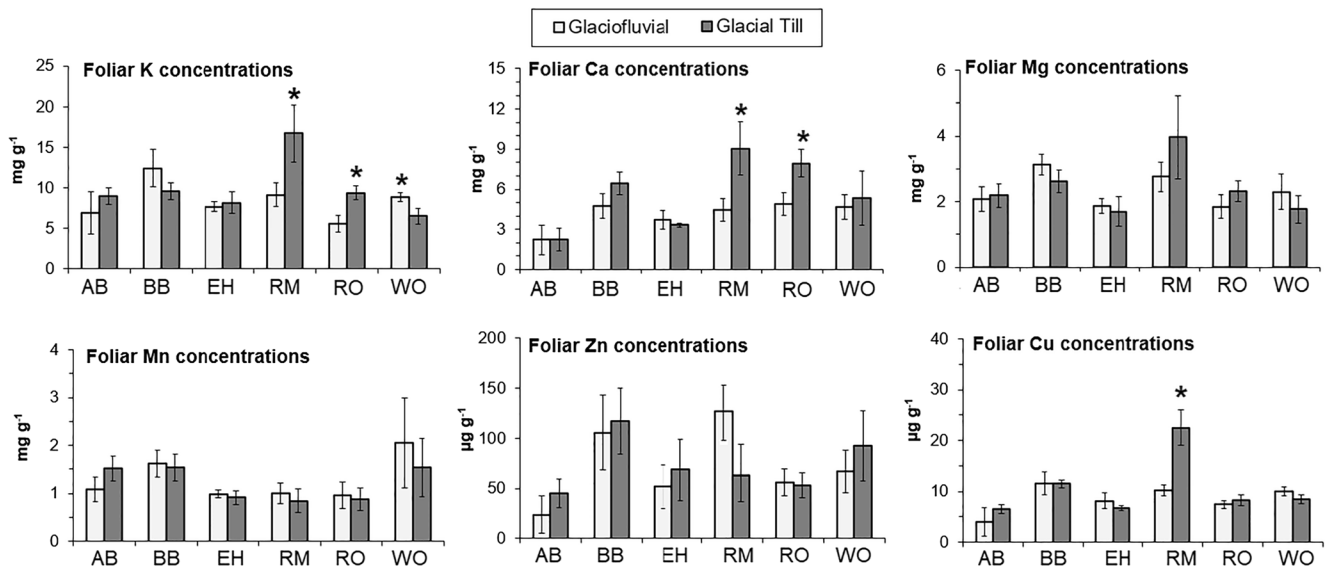
**Fig. 4** Average pseudo-total metal soil concentrations for each soil horizon at glaciofluvial and glacial till sites. Error bars are  $\pm 1$  standard error. (\*) indicates a significant difference between glaciofluvial and glacial till soils



369 significantly different physical and chemical properties.  
 370 Glacial till soils had higher pseudo-total macro- and micronu-  
 371 trient concentrations, but glaciofluvial soils had higher pH and  
 372 finer textures. The finer, well-sorted particle size distribution

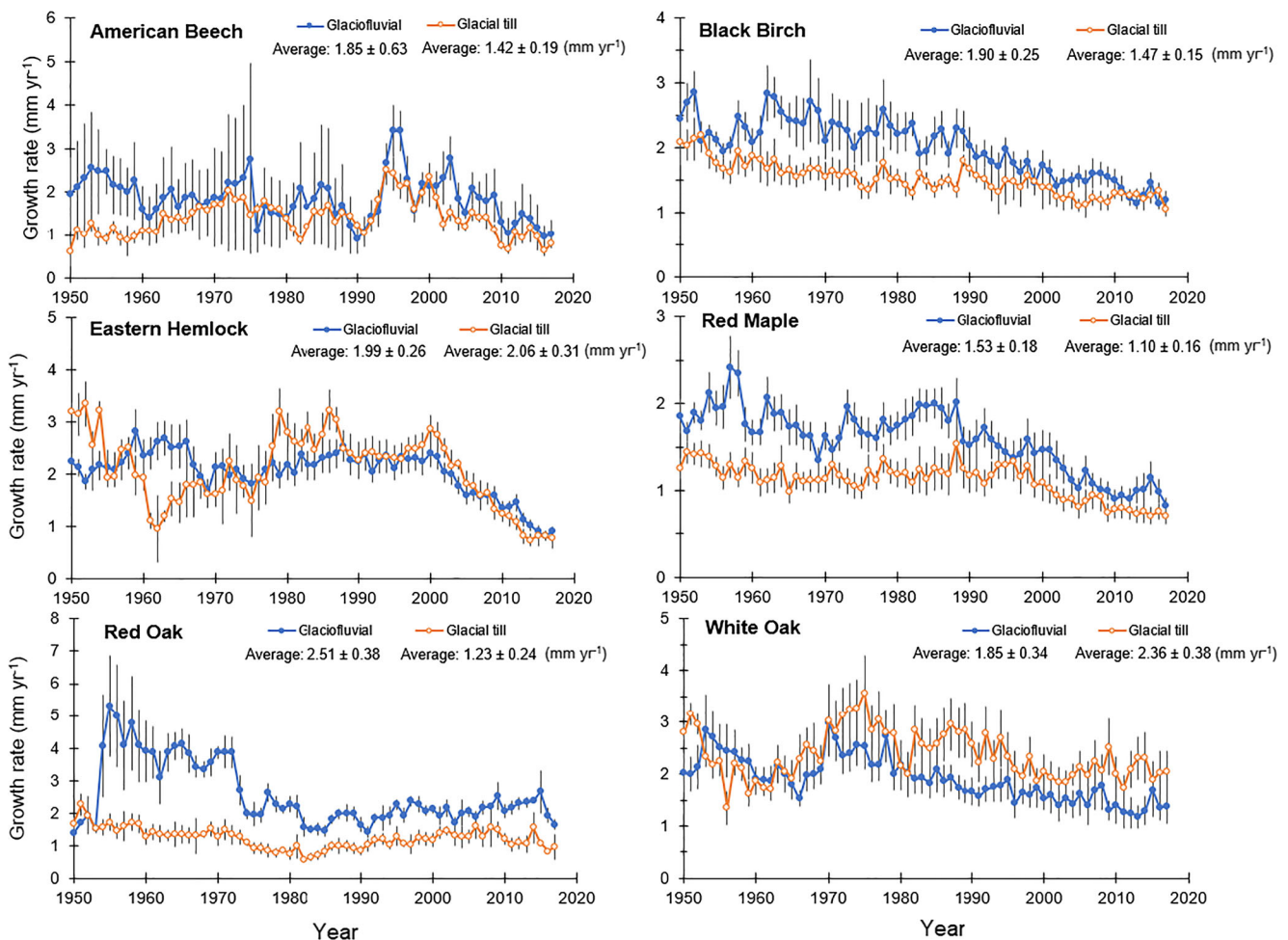
for glaciofluvial soils led to higher water field capacity and  
 less rock fragments, which can decrease water stress in trees  
 during precipitation-limited summer months and increase vol-  
 ume for rooting (Li et al. 2010; Keller and Håkansson 2010;

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**Fig. 5** Average foliar metal concentrations for dominant tree species found across both glaciofluvial and glacial till sites. Error bars are  $\pm 1$  standard error. (\*) indicates a significant difference between glaciofluvial

and glacial till soils. Forest species codes: AB = American Beech, BB = Black Birch, EH = Eastern Hemlock, RM = Red Maple, RO = Red Oak, WO = White Oak



**Fig. 6** Average annual growth rate estimate from tree cores annual ring measurements for dominant tree species found across both glaciofluvial and glacial till sites. Error bars are  $\pm 1$  standard error and  $N=9$  tree cores for each species at across surficial deposit

377 Rab et al. 2011; Olson 2012). Greater clay contents are typi- 378  
 379 cally associated with greater cation exchange and higher sur- 380  
 381 face area for weathering (Miller et al., 1993; Taylor and Blum, 382  
 383 1995), but we observed greater inorganic nutrient concentra- 384  
 385 tions in the sandy glacial till soils. Greater amounts of fine 386  
 387 particles can increase aggregation and stability of C com- 388  
 389 pounds for nutrient retention, particularly in agroforest sys- 390  
 391 tems (e.g., Rocha et al. 2018). The greater inorganic nutrient 392  
 393 concentrations in glacial till soils was likely due to their lower 394  
 395 pH, which agrees with the observations of previous studies 396  
 397 (e.g., Taylor and Blum 1995; Finzi et al. 1998; Nikodemus 398  
 et al. 2013) that greater acidity can increase dissolution and 399  
 leaching of inorganic nutrients from silicates. Further, we hy- 400  
 pothesize that glacial till soils had greater Ca, K, and Mg than 401  
 the glaciofluvial soils because fluvial materials are typically 402  
 more extensively weathered due to reworking by fluvial ac- 403  
 tion. During this weathering and erosional transport, Ca, K, 404  
 and Mg-bearing minerals such as carbonates and apatite are 405  
 lost, leaving behind a greater proportion of resistant, nutrient- 406  
 poor feldspar and quartz (see Harley and Gilkes 2000; Eberl 407  
 2004; Viers et al. 2009). These results support our hypothesis 408

409 that geologic materials control soil properties important for 410  
 tree growth; rocky, glacial soils common in uplands can pro- 411  
 vide greater inorganic nutrients, but low-lying, glaciofluvial 412  
 soils can provide greater water and nutrient retention for 413  
 northern hardwood trees. 414

#### 4.2 Tree Nutrient Uptake and Growth 403

404 Our findings demonstrate that higher pseudo-total Ca, K, Mg, 405  
 Mn, and Zn concentrations in glacial till soils (Fig. 4) did not 406  
 correspond with greater acquisition and uptake of nutrients 407  
 implied through foliar tissue concentrations (Fig. 5) for most 408  
 tree species. These results are novel as, to the authors' knowl- 409  
 edge, this is the first report on geologic material controls on 410  
 northern hardwood nutrient acquisition of macronutrient and 411  
 micronutrient concentrations in temperate forests of New 412  
 England. An analysis of foliar data from the Tree Chemistry 413  
 Database by the U.S. Forest Service (Prado et al. 2002) 414  
 showed that foliar K and Ca concentrations, but not Mg con- 415  
 centrations, may be different when compared among soil par- 416  
 ent materials. The discrepancy between our findings and data



417 from Prado et al. (2002) may be due to a broader climatic  
418 sampling region, wider range of geologic materials included,  
419 or greater number of tree species analyzed. Erdmann et al.  
420 (1988) observed variations in foliar concentrations in Red  
421 Maple across sites but attributed variations to tree physiolog-  
422 ical properties rather than soil properties. Previous research  
423 has primarily focused on N or P cycling in hardwood forests,  
424 but our results show that Red Maple uptake of K, Ca, and Cu  
425 can be affected by geologic material. However, foliar nutrient  
426 concentrations for American Beech, Black Birch, Eastern  
427 Hemlock, Red Oak, and White Oak were similar between  
428 glaciofluvial- and glacial till-derived soils. Acquisition of in-  
429 organic nutrients is essential for chemical signaling, cellular  
430 metabolism, enzyme production, and photosynthesis  
431 (Schaberg et al. 2001; Guo et al. 2016; Zhao et al. 2001;  
432 Wang et al. 2013). One possible mechanism is that soil inor-  
433 ganic nutrient concentrations were adequate for most tree spe-  
434 cies but not low enough to see an effect as observed in the  
435 tropical forests studied by Paoli et al. (2007). An alternative  
436 hypothesis is that most trees were able to acquire similar  
437 amounts of nutrients, regardless of the soil parent material,  
438 due to rhizosphere interactions. As described by Zemunik  
439 et al. (2015), under nutrient-limiting conditions, plants can  
440 adapt for more effective nutrient acquisition through increas-  
441 ing exudate release, stimulating mycorrhizal fungal or bacte-  
442 rial associations, or altering belowground root traits (in  
443 addition, see Uroz et al. 2011; Yin et al. 2014).

444 Our results also demonstrate a non-linear relationship be-  
445 tween soil nutrient concentrations and plant uptake rates, as  
446 represented by foliar concentrations. This could be due to  
447 either the pseudo-total digestion procedure used or the more  
448 likely possibility that trees adapted to increase uptake under  
449 low nutrient availability and limit “luxury” uptake under high  
450 nutrient availability. One possible reason is that pseudo-total  
451 extractions were unable to capture nuances in bioavailability  
452 or type of sorption (such as carbonates, oxide bound, and  
453 organic matter occluded fractions), which can alter the avail-  
454 ability of inorganic nutrients such as Ca and Mg (see Park and  
455 Ro 2018). However, Calvaruso et al. (2017) showed that tree  
456 acquisition and uptake of nutrients are dynamic; trees can  
457 readily adapt to overcome inorganic nutrient constraints in  
458 soils. Thus, we argue that trees obtain nutrients in spite of  
459 lower concentrations in glaciofluvial soils or “luxury uptake”  
460 of nutrients that are limited on glacial till soils. Mineral  
461 weathering of feldspar and apatite has been identified as a  
462 key factor impacting long-term timber harvesting sustainabil-  
463 ity (Vadeboncoeur et al. 2014; Zetterberg et al. 2016). Silicate  
464 minerals can be weathered by secretion of organic compounds  
465 from tree roots (e.g., chelators; see Uroz et al. 2011; Zhu et al.  
466 2014; Yin et al., 2014), or tree-supported microbial commu-  
467 nities may dissolve silicate minerals present (Harley and  
468 Gilkes 2000; Uroz et al. 2009, Ahmed and Holmström  
469 2015). As an example, Zemunik et al. (2015) demonstrated

470 that increased exudation of chelating compounds and stimu-  
471 lation of mycorrhizal fungi increased access of total inorganic  
472 P by plants, not just operationally defined bioavailable P  
473 forms.

474 Lastly, we found the first evidence, to the authors’ knowl-  
475 edge, that geologic materials may control northern hardwood  
476 tree growth in New England. Instead of nutrient limitations,  
477 our data suggests tree growth rates (annual ring thickness  
478  $\text{mm year}^{-1}$ ) were between 1.3 to 2.1 times greater for Black  
479 Birch, Red Maple, and Red Oak on glaciofluvial deposits.  
480 This occurred even though Red Maple and Red Oak had lower  
481 K and Ca soil and foliar concentrations at glaciofluvial sites  
482 than on glacial till sites (Fig. 4). We hypothesize that faster  
483 Black Birch, Red Maple, and Red Oak growth on  
484 glaciofluvial soils than on glacial till was due to soil physical  
485 properties, specifically the significantly greater field capacity,  
486 lower rock fraction, and greater fine fraction (Fig. 3). Previous  
487 literature has focused on light, predation, and diseases as pri-  
488 mary controls on Birch, Maple, and Oak growth rates (e.g.,  
489 Johnson and Abrams, 2009; Parker and Dey 2008).  
490 Kirkpatrick (1981) recognized moisture can control Black  
491 Birch growth but observed their growth across New England  
492 was greater in well-drained, dry soils than on poorly drained,  
493 wet soils. Thus, we demonstrate for the first time that tree  
494 growth for three common northern hardwoods was affected  
495 by the geologic material that served as soil parent material,  
496 which was not related to nutrient uptake or accessibility.

## 5 Conclusions and Implications 497

498 Our study confirmed our hypothesis that geologic materials  
499 can affect tree growth. Black Birch, Red Maple, and Red Oak  
500 were more adept at growing on glaciofluvial geologic deposits  
501 than glacial till. One important implication is that harvesting  
502 common tree species on coarse glacial till materials in western  
503 Massachusetts may affect subsequent tree growth after timber  
504 harvesting. Thus, harvesting Black Birch, Red Maple, and  
505 Red Oak on glacial till may result in slower regeneration while  
506 harvesting these three species on glaciofluvial materials may  
507 result in faster regeneration. Another implication is that tree  
508 species can acquire similar nutrient concentrations, even with  
509 lower available nutrient concentrations. This implies that lim-  
510 itations from mineral weathering and soil retention between  
511 geologic materials can be overcome by mineral–biological  
512 interactions, improving long-term nutrient acquisition. Thus,  
513 operationally defined soil extraction procedures may not ac-  
514 curately capture nutrient availability, particularly when con-  
515 sidering effects from exudates, chelators, and microbial sym-  
516 bionts. Our study only focused on a specific region of western  
517 Massachusetts and Glacial Lake Hitchcock sediments. For  
518 future studies, a greater sampling area across New England  
519 states will greatly enhance the ability to examine if our results

520 are more broadly applicable to and separate from glacial out-  
521 wash, alluvial fans and deltaic deposits and glacial lacustrine  
522 deposits.

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## 525 Compliance with Ethical Standards

526 **Conflict of Interest** The authors declare that they have no conflict of  
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## Q4 528 References

529 Ahmed E, Holmström SJ (2015) Microbe–mineral interactions: the im-  
530 pact of surface attachment on mineral weathering and element se-  
531 lectivity by microorganisms. *Chemical Geol* 403:13–23  
532 Ashley GM (1975) Rhythmic sedimentation in glacial Lake Hitchcock,  
533 Massachusetts–Connecticut. Special publications of Society of  
534 Economic Paleontologists and Mineralogists. Glaciofluvial and  
535 Glaciolacustrine sedimentation (SP23)  
536 Bailey, S.W. and Hornbeck, J.W., 1992. Lithologic composition and rock  
537 weathering potential of forested, glacial-till soils. Res. Pap. NE-662.  
538 Radnor, PA: US. Department of Agriculture, Forest Service, north-  
539 eastern Forest Experiment Station. 7 p., 662  
540 Balco G, Schafer JM (2006) Cosmogenic-nuclide and varve chronolo-  
541 gies for the deglaciation of southern New England. *Quat*  
542 *Geochronol* 1:15–28  
543 Calvaruso C, Kirchen G, Saint-André L, Redon PO, Turpault MP (2017)  
544 Relationship between soil nutritive resources and the growth and  
545 mineral nutrition of a beech (*Fagus sylvatica*) stand along a soil  
546 sequence. *Catena* 155:156–169  
547 Chen M, Ma LQ (1998) Comparison of four USEPA digestion methods  
548 for trace metal analysis using certified and Florida soils. *J Environ*  
549 *Qual* 27:1294–1300  
550 Deal RL, Cochran B, LaRocco G (2012) Bundling of ecosystem services  
551 to increase forestland value and enhance sustainable forest manage-  
552 ment. *For Pol Econ* 17:69–76  
553 Dyke A, Prest V (1987) Late Wisconsinan and Holocene history of the  
554 Laurentide ice sheet. *Géog Phys Quatern* 41:237–263  
555 Dyke AS, Andrews JT, Clark PU, England JH, Miller GH, Shaw J,  
556 Veillette JJ (2002) The Laurentide and Innuitian ice sheets during  
557 the last glacial maximum. *Quat Sci Rev* 21:9–31  
558 Eberl DD (2004) Quantitative mineralogy of the Yukon River system:  
559 changes with reach and season, and determining sediment proven-  
560 nance. *Am Mineral* 89:1784–1794  
561 Erdmann GG, Crow TR, Rauscher HM (1988) Foliar nutrient variation  
562 and sampling intensity for *Acer rubrum* trees. *Can J For Res* 18:  
563 134–139  
564 Finzi AC, Canham CD, Van Breemen N (1998) Canopy tree–soil inter-  
565 actions within temperate forests: species effects on pH and cations.  
566 *Ecol Appl* 8:447–454  
567 Gee GW, Bauder JW (1986) Particle-size analysis 1. Methods of soil  
568 analysis: part 1—physical and mineralogical methods 383–411  
569 Guo W, Nazim H, Liang Z, Yang D (2016) Magnesium deficiency in  
570 plants: an urgent problem. *Crop J* 4:83–91  
571 Hartshorn JH, Young WR (1969) Geography and geology of Glacial  
572 Lake Hitchcock. In: An introduction to the archaeology and history  
573 of the Connecticut Valley Indian, Springfield, Mass, vol 1, pp 19–27  
574 Harley AD, Gilkes RJ (2000) Factors influencing the release of plant  
575 nutrient elements from silicate rock powders: a geochemical over-  
576 view. *Nutr Cycl Agroecosyst* 56:11–36

Johnson SE, Abrams MD (2009) Age class, longevity and growth rate  
relationships: protracted growth increases in old trees in the eastern  
United States. *Tree Physiol* 29:1317–1328 577  
578  
579  
Joshi O, Mehmood SR (2011) Factors affecting nonindustrial private  
forest landowners' willingness to supply woody biomass for  
bioenergy. *Biomass Bioenergy* 35:186–192 580  
581  
582  
Keller T, Håkansson I (2010) Estimation of reference bulk density from  
soil particle size distribution and soil organic matter content.  
*Geoderma* 154:398–406 583  
584  
585  
Kirkpatrick M (1981) Spatial and age dependent patterns of growth in  
New England black birch. *Am J Bot* 68:535–543 586  
587  
Legout A, Hansson K, Van Der Heijden G, Laclau J-P, Augusto L,  
Ranger J (2014) Chemical fertility of forest soils: basic concepts.  
*Revue forestière française* 66:21–31 588  
589  
590  
Li Z, Wu P, Feng H, Zhao X, Huang J, Zhuang W (2010) Simulated  
experiment on effects of soil bulk density on soil water holding  
capacity. *Acta Pedol Sin* 47:611–620 591  
592  
593  
Li D, Wen L, Zhang W, Yang L, Xiao K, Chen H, Wang K (2017)  
Afforestation effects on soil organic carbon and nitrogen pools mod-  
ulated by lithology. *For Ecol Manag* 400:85–92 594  
595  
596  
Lilliefors HW (1967) On the Kolmogorov–Smirnov test for normality  
with mean and variance unknown. *J Am Stat Assoc* 62:399–402 597  
598  
Miller EK, Blum JD, Friedland AJ (1993) Determination of soil  
exchangeable-cation loss and weathering rates using Sr isotopes.  
*Nature* 362:438 599  
600  
601  
Mohanty BP, Mousli Z (2000) Saturated hydraulic conductivity and soil  
water retention properties across a soil-slope transition. *Water*  
*Resour Res* 36:3311–3324 602  
603  
604  
Nelson G, Earle CJ, Spellenberg R (2014) Trees of eastern North  
America, vol 93. Princeton University Press 605  
606  
Nikodemus O, Kasparinskis R, Kukuls I (2013) Influence of afforestation  
on soil genesis, morphology and properties in glacial till deposits.  
*Arch Agron Soil Sci* 59:449–465 607  
608  
609  
Olson G (2012) Soils and the environment: a guide to soil surveys and  
their applications. Springer Science & Business Media 610  
611  
Paoli GD, Curran LM, Slik JW (2007) Soil nutrients affect spatial patterns  
of aboveground biomass and emergent tree density in southwestern  
Borneo. *Oecologia* 155:287–299 612  
613  
614  
Park JS, Ro HM (2018) Early-stage changes in chemical phosphorus  
speciation induced by liming deforested soils. *J Soil Sci Plant Nutr*  
18:435–447 615  
616  
617  
Parker WC, Dey DC (2008) Influence of overstorey density on ecophys-  
iology of red oak (*Quercus rubra*) and sugar maple (*Acer*  
*saccharum*) seedlings in Central Ontario shelterwoods. *Tree*  
*Physiol* 28:797–804 618  
619  
620  
621  
Pardo, Linda H.; Robin-Abbott, Molly; Duarte, Natasha; Miller, Eric K.  
2005. Tree chemistry database (version 1.0). Gen. Tech. Rep. NE-  
324. Newtown Square PA: U.S. Department of Agriculture, Forest  
Service, Northeastern Research Station. 45 p 622  
623  
624  
625  
Paré D, Thiffault E (2016) Nutrient budgets in forests under increased  
biomass harvesting scenarios. *Curr For Rep* 2:81–91 626  
627  
628  
629  
630  
631  
632  
633  
634  
635  
636  
637  
638  
639  
640  
Rab MA, Chandra S, Fisher PD, Robinson NJ, Kitching M, Aumann CD,  
Imhof M (2011) Modelling and prediction of soil water contents at  
field capacity and permanent wilting point of dryland cropping soils.  
*Soil Res* 49:389–407  
Rawls WJ, Gish TJ, Brakensiek DL (1991) Estimating soil water reten-  
tion from soil physical properties and characteristics. In *Advances in*  
*soil science*. Pg 213–234. Springer, New York  
Rayburn JA, Vollmer FW (2013) ANTEVS: a quantitative varve se-  
quence cross-correlation technique with examples from the north-  
eastern USA. *GFF* 135:282–292  
Rechcigl JE, Payne GG (1990) Comparison of a microwave digestion  
system to other digestion methods for plant tissue analysis.  
*Commun Soil Sci Plant Anal* 21(19–20):2209–2218

- 641 Richardson JB, Friedland AJ (2016) Influence of coniferous and deciduous  
642 vegetation on major and trace metals in forests of northern New  
643 England, USA. *Plant Soil* 402:363–378
- 644 Ridge JC, Larsen FD (1990) Re-evaluation of Antevs' New England  
645 varve chronology and new radiocarbon dates of sediments from  
646 glacial Lake Hitchcock. *Geol Soc Am Bull* 102:889–899
- 647 Robertson G, Gualke P, McWilliams R, LaPlante S, Guldin R (2011)  
648 National report on sustainable forests–2010. USDA Forest  
649 Service, Washington, pp FS, 212 pp–979
- 650 Rocha PRD Jr, Ribeiro PH, Mesquita LF, Andrade FV, Mendonça EDS  
651 (2018) Distribution of C and inorganic phosphorus fractions in dif-  
652 ferent aggregate sizes under forestry, agroforestry system and pas-  
653 ture. *J Soil Sci Plant Nutr* 18:361–375
- 654 Royer-Tardif S, Bradley RL (2011) Evidence that soil fertility controls the  
655 mixing of jack pine with trembling aspen. *For Ecol Manag* 262:  
656 1054–1060
- 657 Schaberg PG, DeHayes DH, Hawley GJ (2001) Anthropogenic calcium  
658 depletion: a unique threat to forest ecosystem health? *Ecosyst Health*  
659 7:214–228
- 660 Schoeneberger PJ, Wysocki DA, Benham EC (2012) Field book for de-  
661 scribing and sampling soils. Natural Resources Conservation  
662 Service, National Soil Survey Center, Lincoln
- 663 Soil Survey Staff (2008) Natural Resources Conservation Service, United  
664 States Department of Agriculture. Web soil survey. Available online  
665 at the following link: <https://websoilsurvey.sc.egov.usda.gov/>.  
666 Accessed [06/21/2018]
- 667 Taylor A, Blum JD (1995) Relation between soil age and silicate  
668 weathering rates determined from the chemical evolution of a glacial  
669 chronosequence. *Geology* 23(11):979–982
- 670 Uchupi E, Driscoll N, Ballard RD, Bolmer ST (2001) Drainage of late  
671 Wisconsin glacial lakes and the morphology and late quaternary  
672 stratigraphy of the New Jersey–southern New England continental  
673 shelf and slope. *Mar Geol* 172(1–2):117–145
- 674 Uroz S, Calvaruso C, Turpault MP, Frey-Klett P (2009) Mineral  
675 weathering by bacteria: ecology, actors and mechanisms. *Trends*  
676 *Microbiol* 17:378–387
- 677 Uroz S, Oger P, Lepleux C, Collignon C, Frey-Klett P, Turpault MP  
678 (2011) Bacterial weathering and its contribution to nutrient cycling  
679 in temperate forest ecosystems. *Res Microbiol* 162:820–831
- 680 Viers J, Dupré B, Gaillardet J (2009) Chemical composition of suspended  
681 sediments in world rivers: new insights from a new database. *Sci*  
682 *Total Environ* 407:853–868
- 683 Wang M, Zheng Q, Shen Q, Guo S (2013) The critical role of potassium  
684 in plant stress response. *Int J Mol Sci* 14:7370–7390
- 685 Watabe Y, Leroueil S, Le Bihan JP (2000) Influence of compaction con-  
686 ditions on pore-size distribution and saturated hydraulic conductiv-  
687 ity of a glacial till. *Can Geotech J* 37:1184–1194
- 688 Vadeboncoeur MA, Hamburg SP, Yanai RD, Blum JD (2014) Rates of  
689 sustainable forest harvest depend on rotation length and weathering  
690 of soil minerals. *For Ecol Manag* 318:194–205
- 691 Villholth KG, Jensen KH, Fredericia J (1998) Flow and transport pro-  
692 cesses in a macroporous subsurface-drained glacial till soil I: field  
693 investigations. *J Hydrol* 207:98–120
- 694 Yin H, Wheeler E, Phillips RP (2014) Root-induced changes in nutrient  
695 cycling in forests depend on exudation rates. *Soil Biol Biochem* 78:  
696 213–221
- 697 Zetterberg T, Olsson BA, Löfgren S, Hyvönen R, Brandtberg PO (2016)  
698 Long-term soil calcium depletion after conventional and whole-tree  
699 harvest. *For Ecol Manag* 369:102–115
- 700 Zemunik G, Turner BL, Lambers H, Laliberté E (2015) Diversity of plant  
701 nutrient-acquisition strategies increases during long-term ecosystem  
702 development. *Nat Plants* 1:15050
- 703 Zhao D, Oosterhuis DM, Bednarz CW (2001) Influence of potassium  
704 deficiency on photosynthesis, chlorophyll content, and chloroplast  
705 ultrastructure of cotton plants. *Photosynthetica* 39:103–109
- 706 Zhu Y, Duan G, Chen B, Peng X, Chen Z, Sun G (2014) Mineral  
707 weathering and element cycling in soil-microorganism-plant sys-  
708 tem. *Sci China Earth Sci* 57:888–896
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tional claims in published maps and institutional affiliations. 710
- 711